

A Global 3-D Model Evaluation of the Atmospheric Budgets of HCN and CH₃CN: Constraints From Aircraft Measurements Over the Western Pacific

Qinbin Li, Daniel J. Jacob, Robert M. Yantosca, Colette L. Heald
Department of Earth and Planetary Sciences and Division of Engineering and Applied
Sciences, Harvard University, Cambridge, MA

Hanwant B. Singh
NASA Ames Research Center, Moffett Field, CA

Makoto Koike
Department of Earth and Planetary Science, University of Tokyo, Tokyo, Japan

Yongjing Zhao
Department of Physics, University of Toronto, Toronto, Canada.

Glen W. Sachse
NASA Langley Research Center, Hampton, VA

David G. Streets
Argonne National Laboratory, Argonne, IL

Short title: ATMOSPHERIC BUDGETS OF HCN AND CH₃CN

Abstract.

We construct global atmospheric budgets of HCN and CH₃CN through a 3-D simulation of the HCN-CH₃CN-CO system constrained and evaluated with aircraft observations from the TRACE-P mission over the NW Pacific in February-April 2001. Observed background vertical gradients of HCN and CH₃CN imply a dominant ocean sink for both gases, with deposition velocity of 0.13 cm s⁻¹ for both and saturation ratios of 0.79 for HCN and 0.88 for CH₃CN. Observations for both gases in the free troposphere implied a dominant source from biomass burning. Enhancement of HCN observed in Chinese urban plumes is attributed tentatively to residential coal burning. Biomass burning and residential coal burning emission ratios relative to CO of 0.27% and 1.6% respectively for HCN, and of 0.20% and 0.25% respectively for CH₃CN, are consistent with observations in biomass burning and Chinese urban plumes, and provide the best fit in the model for simulation of observed TRACE-P vertical profiles, HCN-CH₃CN-CO correlations, as well as long-term records of HCN columns and CH₃CN observations over the northern Indian Ocean. Biomass burning and residential coal burning contribute 0.63 and 0.2 Tg N yr⁻¹ respectively to global HCN and 0.47 and 0.03 Tg N yr⁻¹ respectively to CH₃CN. Ocean uptake is the dominant sink for both gases, with oxidation by OH representing an additional minor sink. The resulting tropospheric lifetimes are 5.3 months for HCN and 5.8 months for CH₃CN. The model predicts very low HCN and CH₃CN concentrations at high southern latitudes, reflecting the assumption of uniform saturation ratio; observations in that region are needed. In the free troposphere, the dominance of biomass burning sources (70-85% for HCN and 90-95% for CH₃CN) implies that both gases can be used as biomass burning tracers. More work is needed to identify the urban source apparent in the Chinese plume observations.

1. Introduction

Hydrogen cyanide (HCN) and methyl cyanide (CH₃CN, also called acetonitrile) are atmospheric tracers of biomass burning [Lobert *et al.*, 1990; Holzinger *et al.*, 1999] and could play a non-negligible role in the nitrogen cycle [Li *et al.*, 2000]. Ocean uptake has been hypothesized as their dominant sink, with corresponding lifetimes of a few months [Hamm and Warneck, 1990; Li *et al.*, 2000]. However, their atmospheric budgets are still poorly understood. We use here a global 3-D model analysis of HCN and CH₃CN observations in Asian outflow from the TRACE-P aircraft mission over the NW Pacific [Singh *et al.*, this issue] to improve the constraints on the sources and sinks of these two gases. The TRACE-P data represent the first in situ measurements for HCN in the troposphere, and one of the first for CH₃CN. Previous data for HCN are from remote sensing of the total column at several sites around the world [Mahieu *et al.*, 1995, 1997; Rinsland *et al.*, 1999, 2000, 2001, 2002; Zhao *et al.*, 2000, 2002]. Previous in situ measurements of CH₃CN have been made over the Indian Ocean [de Laat *et al.*, 2001] and over the Amazon [Williams *et al.*, 2001].

It is well established from laboratory and field experiments that biomass burning is a major source for both HCN (0.1-3.2 TgNyr⁻¹) and CH₃CN (0.2-1.1 TgNyr⁻¹) (Table 1). It is also well established that sources from automobile exhaust and industrial processes are negligible in comparison [Lobert *et al.*, 1991; Bange, 2000; Holzinger *et al.*, 2001]. A recent field experiment indicates no emission of HCN from domestic biofuels [Bertschi *et al.*, 2002]. Atmospheric sinks of HCN and CH₃CN from reaction with OH and O(¹D), photolysis, and scavenging by precipitation yield lifetimes of a few years [Cicerone and Zellner, 1983; Brasseur *et al.*, 1983]. The solubilities of HCN and CH₃CN are sufficiently high for ocean uptake to impose atmospheric lifetimes of a few months if loss in the oceanic mixed layer is sufficiently rapid [Hamm and Warneck, 1990; Li *et al.*, 2000]. There is some evidence that HCN and CH₃CN are consumed biologically in the ocean [Singh *et al.*, this issue]. Li *et al.* [2000] showed in a global 3-D model study that the observed seasonal variation of the HCN column in different regions of the world is consistent with a scenario where biomass burning provides the main source and ocean uptake provides the main sink. They inferred a global HCN biomass burning source of 1.4-2.9 TgNyr⁻¹, and an oceanic saturation ratio of 0.83 or less. Observed vertical gradients of HCN and CH₃CN in remote marine air in TRACE-P confirm the importance of the ocean sink [Singh *et al.*, this issue].

The TRACE-P mission was conducted in February-April 2001 over the NW Pacific to study the transport and chemical evolution of Asian outflow over the Pacific [Jacob *et al.*, this

issue]. It used two NASA aircraft, a DC-8 and a P-3B, operating out of Hong Kong and Japan, and with additional transit flights and sorties over the North Pacific. Biomass burning in SE Asia is an important component of Asian outflow in spring [*Liu et al.*, 1999; *Chan et al.*, 2000; *Bey et al.*, 2001b]. During TRACE-P, Asian biomass burning and anthropogenic sources had contributions of similar magnitude to CO concentrations observed in Asian outflow [*Liu et al.*, this issue; *Heald et al.*, this issue].

Both HCN and CH₃CN were measured by gas chromatography aboard the DC-8 aircraft [*Singh et al.*, this issue]. The measurement accuracy was about $\pm 25\%$ and the detection limit 30 pptv. Other measurements aboard the DC-8 included CO, perchloroethylene (C₂Cl₄), and methyl chloride (CH₃Cl). CO is a tracer of combustion. C₂Cl₄, a synthetic organic chemical used in dry cleaning, is an anthropogenic pollution tracer. CH₃Cl can be used as a biomass burning tracer in Asian outflow [*Blake et al.*, 2001]. These complementing measurements can be used to place constraints on the sources of HCN and CH₃CN [*Singh et al.*, this issue]. Some of the results shown in *Singh et al.* [this issue] are reproduced here for purpose of comparison to model results. Throughout this paper, model comparisons to observations will use model results sampled along the DC-8 flight paths. All linear regressions shown in this paper are calculated using the reduced major axis (RMA) method, which allows for errors in both variables.

2. Model Simulation

We simulate the distributions of HCN, CH₃CN, and CO observed in TRACE-P with the GEOS-CHEM global 3-D model of tropospheric chemistry. A detailed description of the model is presented by *Bey et al.* [2001a]. We use here GEOS-CHEM version 4.33 (see <http://www-as.harvard.edu/chemistry/trop/geos>). The model is driven by assimilated meteorological fields from the Goddard Earth Observing System (GEOS) of the NASA Data Assimilation Office (DAO). We use meteorological fields for 2000-2001 (GEOS-3) which are provided with 6-hour temporal resolution (3-hour for surface variables and mixed layer depths), $1^\circ \times 1^\circ$ horizontal resolution, and 48 vertical sigma levels. We degrade the horizontal resolution to $2^\circ \times 2.5^\circ$ for application in GEOS-CHEM. The simulations are conducted for a 16-month period (January 2000-April 2001). The first 13 months are used for initialization and we present results for the February-April 2001 TRACE-P period.

Sources and sinks of HCN and CH₃CN in the model are described in section 3. The

CO simulation follows that of *Duncan et al.* [2002] and uses archived OH concentration fields (monthly mean) from a full-chemistry simulation [*Li et al.*, 2002] to calculate CO loss. Anthropogenic emissions are as described in *Duncan et al.* [2002] except for Asia, where the *Streets et al.* [this issue] inventory is used. Biomass burning emissions are as described in *Heald et al.* [this issue]. The total biomass burning CO source is 487 TgCOyr^{-1} for 2000, while the total CO sources from fossil fuel and biofuel are 497 and 200 TgCOyr^{-1} , respectively. For the TRACE-P period (February-April 2001), the global biomass burning CO source is 129 TgCO and about half of that source is in SE Asia.

Since the model simulations of HCN, CH_3CN , and CO as implemented here are linear with respect to their sources, we resolve the contributions from different source regions and source types in the model by using tagged tracers [*Bey et al.*, 2001b], which can then be added to reconstitute the full concentration field. Tagged tracers resolve sources from biomass burning, biofuels, and residential coal burning, for Asia and for the rest of the world.

3. Model Sources and Sinks of HCN and CH_3CN

3.1. Sinks

3.1.1. Ocean Uptake Constraints on ocean uptake for HCN and CH_3CN can be obtained from the vertical gradients of concentrations measured in background air over the North Pacific during TRACE-P [*Singh et al.*, this issue]. We implement this constraint here by using the observed gradients to infer a saturation ratio for these two gases that we can then extrapolate globally in the 3-D model simulation.

Figure 1 shows the vertical distributions of HCN, CH_3CN , and CO observed for the ensemble of the data in TRACE-P (dots and red lines). The CO mixing ratios decrease from the boundary layer to the free troposphere, while the opposite is seen in the mixing ratios of HCN and CH_3CN . This is consistent with an ocean sink for HCN and CH_3CN but could also reflect the dominant biomass burning source for these two gases, in contrast to the large fossil fuel source for CO. Biomass burning outflow during TRACE-P was largely confined to the free troposphere, while there was considerable fossil fuel outflow in the boundary layer [*Carmichael et al.*, this issue; *Liu et al.*, this issue].

Figure 2 shows the vertical distributions of HCN and CH_3CN (dots and red lines) observed in TRACE-P over the North Pacific under background conditions, defined here as $\text{CO} < 120 \text{ ppbv}$ and $\text{C}_2\text{Cl}_4 < 10 \text{ pptv}$ [*Singh et al.*, this issue]. Both gases show significantly

reduced mixing ratios in the marine boundary layer compared with their free tropospheric values. Asian outflow influence for these conditions is minimal, and the shape of the profiles is evidence for ocean uptake. Median mixing ratios at 2-4 km are 201 pptv for HCN and 134 pptv for CH₃CN, while the corresponding values at 0-2 km are 161 pptv and 106 pptv, respectively.

Following *Singh et al.* [this issue], the deposition velocities of HCN and CH₃CN to the ocean can be derived using a box model of the marine boundary layer (MBL) with no divergence of horizontal flux [*Lenschow*, 1983]. We calculate the flux across the MBL top as the concentration difference between 2-4 km (C_{FT}) and 0-2 km (C_{MBL}) multiplied by an entrainment velocity $K_e = 0.5 \text{ cm s}^{-1}$ [*Heikes et al.*, 1996]. At steady state, this flux must be balanced by the ocean uptake flux ($v_d C_{MBL}$, where v_d is the deposition velocity). Loss by reaction with OH is negligible, since both HCN and CH₃CN can be considered chemically inert on the time scale for MBL turnover (a few days). It follows that

$$v_d = K_e \left(\frac{C_{FT}}{C_{MBL}} - 1 \right) \quad (1)$$

The resulting deposition velocity is 0.13 cm s^{-1} for both HCN and CH₃CN.

The comparable deposition velocities despite greater solubility of CH₃CN imply faster consumption of HCN/CN⁻ in the oceans (see discussion below). *de Laat et al.* [2001] previously inferred a deposition velocity of $0.01\text{-}0.05 \text{ cm s}^{-1}$ for CH₃CN to the ocean in order to explain CH₃CN mixing ratios observed in South Asian outflow to the Indian Ocean (INDOEX mission). *Warneke and de Gouw* [2001] estimated a deposition velocity of 0.34 cm s^{-1} for CH₃CN from ship measurements taken during an oceanic upwelling event. They attributed the higher value compared to *de Laat et al.* [2001] to the lower surface temperature and higher biological activity in the upwelling water. More recently, *de Gouw et al.* [2002] estimated a deposition velocity of 0.17 cm s^{-1} for CH₃CN from aircraft measurements over the NE Pacific.

Using a two-layer film model for air-to-sea exchange [*Liss and Slater*, 1974], *Singh et al.* [this issue] estimated saturation ratios of 0.77 for HCN and 0.83 for CH₃CN from their TRACE-P data. Saturation ratio is related to deposition velocity by

$$v_d = (1 - S) K_g \quad (2)$$

where K_g is the air-to-sea transfer velocity and is calculated from conductances for mass transfer in the gas and liquid phases, for which we adopt the *Asher* [1997] and the *Nightingale*

et al. [2000] parameterizations, respectively. The resulting saturation ratios are 0.79 for HCN and 0.88 for CH₃CN. The difference with *Singh et al.* [this issue] reflects different assumptions on the parameterizations of conductances for mass transfer. In our 3-D model calculation we extrapolate these saturation ratios globally, resulting in atmospheric lifetimes against ocean uptake of six months for HCN and 8 months for CH₃CN.

At steady state, the ocean uptake flux must be balanced by consumption in the oceanic mixed layer (OML)

$$v_d C_{MBL} = \frac{Z}{\tau} C_{OML} \quad (3)$$

$$S = H \frac{C_{OML}}{C_{MBL}} \quad (4)$$

$$\tau = \frac{Z}{H} \frac{S}{v_d} = \frac{Z}{H} \left(\frac{1}{v_d} - \frac{1}{K_g} \right) \quad (5)$$

where Z is the oceanic mixed layer depth (about 50 m for the middle-latitude North Pacific in March [*Kara et al.*, 2000]), τ is the lifetime against consumption in the OML, C_{OML} is the concentration in the OML, and H is the dimensionless Henry's law constant defined as the ratio of the concentration in air to that in seawater. Values for H at 298 K are 3.4×10^{-3} ($\Delta H_{298}/R = 5000$ K) for HCN and 7.6×10^{-4} ($\Delta H_{298}/R = 4100$ K) for CH₃CN [*Sander*, 1999]. The resulting lifetimes against consumption in the OML are three months for HCN(aq)/CN⁻ and 14 months for CH₃CN(aq)/CN⁻.

3.1.2. Other Sinks Additional known losses for atmospheric HCN and CH₃CN are reactions with OH and O(¹D), photolysis, and precipitation scavenging. Of these, only reaction with OH is significant [*Cicerone and Zellner*, 1983; *Brasseur et al.*, 1983] and is implemented in the model using rate constants from *Wine* [2002] for HCN and *DeMore et al.* [1997] for CH₃CN, with monthly mean OH fields from a full chemistry simulation [*Li et al.*, 2002]. The recent laboratory study of *Wine* [2002] indicates a factor of three lower rate constant for the HCN-OH reaction (low-pressure limit 7.4×10^{-33} cm⁶ molecule⁻² s⁻¹; high-pressure limit 9.0×10^{-15} (T/300)^{3.2} cm³ molecule⁻¹ s⁻¹, where T is temperature) than previously measured [*Fritz et al.*, 1984]. The corresponding lifetimes against oxidation by OH are 4.3 years for HCN and 2.2 years for CH₃CN, much longer than the lifetimes against ocean uptake.

3.2. Sources

3.2.1. Biomass Burning Both HCN and CH₃CN are mainly emitted in the low-temperature smoldering phase of biomass burning, similarly to CO [*Lobert et al.*, 1990].

Different fuel types and fire temperatures result in different emission ratios [Lobert *et al.*, 1991]. Past studies have indicated a large range of molar emission ratios for HCN (0.03-1.1%) and a narrower range for CH₃CN (0.12-0.25%) [Lobert *et al.*, 1991; Hurst *et al.*, 1994a, b; Yokelson *et al.*, 1997, 1999, 2002; Holzinger *et al.*, 1999; Goode *et al.*, 2000]. Emission ratios throughout this paper are in molar units relative to CO. When compounded with a range of global biomass burning CO emission estimates (200-800 Tg CO yr⁻¹), one obtains global biomass burning sources of 0.1-3.2 Tg N yr⁻¹ for HCN and 0.1-1.1 Tg N yr⁻¹ for CH₃CN (Table 1).

The recent review of Andreae and Merlet [2001] recommends biomass burning HCN emission ratios of 0.05% (savanna and grassland) and 0.15% (tropical and extratropical forests) and a biomass burning CH₃CN emission ratio of 0.11% for all vegetation types. The HCN recommendation for savanna and grassland is based on field experiments in Australia [Hurst *et al.*, 1994a, b]. Recent measurements in African savanna fires show much higher HCN emission ratios ($0.85 \pm 0.29\%$) [Yokelson *et al.*, 2002]. By averaging measurements of biomass burning emissions from different fuel types and from different regions of the world, Goode *et al.* [2000] found an average biomass burning HCN emission ratio of 0.34%, much higher than the Andreae and Merlet [2001] value.

Correlations with CO in the TRACE-P observations provide important constraints on the emissions of HCN and CH₃CN. Singh *et al.* [this issue] used these correlations in biomass burning plumes to derive emission ratios of $0.27 \pm 0.12\%$ for HCN and $0.15 \pm 0.05\%$ for CH₃CN. The HCN emission ratio is consistent with the average value reported by Goode *et al.* [2000], and is also in accord with the value of 0.28% estimated by Heald *et al.* [this issue] based on multivariate analysis of the TRACE-P CO and HCN data. The CH₃CN emission ratio is consistent with the Andreae and Merlet [2001] value. We find in GEOS-CHEM that the best fit to the TRACE-P observations of HCN, CH₃CN, and CO, and to the global HCN column data is obtained with biomass burning emission ratios of 0.27% for HCN and 0.20% for CH₃CN, consistent with Singh *et al.* [this issue]; we use these values in the simulations presented here. The corresponding sources of HCN and CH₃CN are 0.63 and 0.47 Tg N yr⁻¹, respectively (Table 2), within the range of past estimates (Table 1).

3.2.2. Chinese Urban Source It is well established that HCN and CH₃CN are produced in the early stages of high-temperature combustion but are subsequently oxidized within the combustion chamber [Flagan and Seinfeld, 1988]. Direct measurements in automobile exhaust [Lobert *et al.*, 1991; Holzinger *et al.*, 2001] as well as aircraft measurements by Singh

et al. [this issue] and *de Gouw et al.* [2002] over southern California confirm that this source is negligible compared to biomass burning (Table 1) and we do not include it in the present study.

However, TRACE-P observations showed a strong enhancement of HCN, and a much weaker enhancement of CH₃CN, in Chinese urban plumes sampled in the MBL less than two days downwind of the source [*Singh et al.*, this issue]. Figure 3 shows the HCN and CH₃CN correlations with CO observed in three major Chinese urban plumes (DC-8 flights 9, 12, 13) sampled in the MBL during TRACE-P. Asian outflow in the MBL was in general advected behind cold fronts and was devoid of biomass burning influence [*Carmichael et al.*, this issue; *Liu et al.*, this issue]. The peak mixing ratios of HCN and CO in the urban plumes are 1.5 ppbv and 1 ppmv, respectively, while the peak mixing ratio for CH₃CN is 300 pptv, much less enhanced. These peak mixing ratios were associated with high mixing ratios of CH₃Cl (see Figure 7a of *Singh et al.* [this issue]) and carbonyl sulfide (COS). The HCN-CO and CH₃CN-CO correlations show slopes of 0.17% and 0.04%, respectively (Figure 3).

Although there is widespread biofuel use in China [*Streets et al.*, this issue], data from Africa indicate negligible nitrile emission from biofuels [*Bertschi et al.*, 2002; *Yokelson et al.*, 2002]. Also, the CH₃CN/HCN enhancement ratio is much lower in the Chinese urban plumes than in biomass burning plumes (Figure 4). *Singh et al.* [this issue] suggested that the enhancements of CH₃Cl and COS correlated with enhanced HCN in the Chinese urban plumes could indicate a source from hard coal burning, and we follow that suggestion here. Clearly, further work is needed identify the Chinese source of nitriles.

The *Streets et al.* [this issue] inventory includes explicitly residential coal burning source for Asia. For the rest of the world, this source is taken from the EDGAR 3.2 emission inventory [*Olivier and Berdowski*, 2001]. In *Duncan et al.* [2002], residential coal burning was lumped into fossil fuel source. We did not bother to separate it out because it is only a small fraction of the total anthropogenic source. The resulting CO source from residential coal burning is 30 Tg CO yr⁻¹ in the model.

We estimate the residential coal burning emission ratios of HCN and CH₃CN from the three Chinese urban plumes observed in TRACE-P (Figure 3). According to the *Streets et al.* [this issue] emission inventory, emissions from residential coal burning account for 7% of the total CO emissions for China overall. Combined with the slopes of the HCN-CO and CH₃CN-CO correlations observed in the Chinese urban plumes (Figure 3), we obtain residential coal burning emission ratios of 2.4% for HCN and 0.6% for CH₃CN. Simulation

of the TRACE-P data with GEOS-CHEM (section 4) indicates a best fit for residential coal burning emission ratios of 1.6% for HCN and 0.25% for CH₃CN. Additionally, we find in the GEOS-CHEM that to achieve the best fit to the observed HCN-CH₃CN-CO correlations in the boundary layer, we have to double the residential coal burning source from Asia. This leads us to reduce the residential coal burning source from the rest of the world by a factor of eight to achieve a best fit to the observed vertical distributions of HCN and CH₃CN.

When scaling the above emission ratios by the fraction of urban CO from residential coal burning in China, the resulting HCN/CO source ratio (0.24% for Shanghai, for example) is comparable to biomass burning (0.27%) while the resulting CH₃CN/CO source ratio (0.04% for Shanghai, for example) is much lower than biomass burning (0.20%). The corresponding global sources are 0.2 for HCN and 0.03 TgNyr⁻¹ for CH₃CN (Table 2). This source could have a seasonal variation associated with heating [*Streets et al.*, this issue], but uncertainties are sufficiently large that we do not bother to include it here.

3.2.3. Biogenic Source There is indication that HCN is produced by higher plants [see *Cicerone and Zellner*, 1983, and references therein]. Recent laboratory results by *Fall et al.* [2001] show that wounded clover releases HCN in addition to large amounts of acetone. In the TRACE-P data, HCN is correlated with both acetone ($r = 0.81$, slope = 0.33 pptv ppbv⁻¹, $n = 746$) and methanol ($r = 0.86$, slope = 136.6 pptv ppbv⁻¹, $n = 744$), which have strong biogenic sources. In particular, methanol is generally considered a tracer for continental biogenic emissions [*Heikes et al.*, 2002]. Figure 5 (top panel) shows the correlation of HCN with methanol observed during TRACE-P. A similarly strong correlation was found for CH₃CN. The biogenic source of methanol is estimated to be 50-280 Tg yr⁻¹ [*Heikes et al.*, 2002]. If the HCN-methanol correlation were to reflect a biogenic source, then the implied biogenic HCN source would be 3-17 TgNyr⁻¹, far too large to be reconciled with the observed HCN concentrations [*Li et al.*, 2000]. Additionally, the highest methanol and HCN concentrations in TRACE-P were observed in the Chinese urban plumes. Methanol in TRACE-P was also strongly correlated with CO ($r = 0.73$, slope = 81.9 ppbv ppbv⁻¹, $n = 1034$) and was highest (together with HCN) in the Chinese urban plumes (Figure 5, bottom panel), implying an anthropogenic source. Thus we find no convincing evidence of a biogenic source of nitriles and do not include such a source in our simulations.

4. Simulation of Observations

We examine in this section how the global specification of sources and sinks of HCN and CH₃CN presented in section 3 allows a successful simulation of the TRACE-P observations and what constraints this successful simulation provides. The model evaluation also considers CO, since our source specifications for HCN and CH₃CN are based on relationships with CO. We also examine the consistency of our simulation with the global data sets of HCN columns previously used as constraints by *Li et al.* [2000] and with the INDOEX data of *de Laat et al.* [2001].

As a general assessment of our simulation, we show in Figure 6 the frequency distributions of the observed and simulated HCN, CH₃CN, and CO concentrations along the ensemble of TRACE-P flight tracks. The simulated mean and median values agree with the observed values to within 10 pptv for HCN, 5 pptv for CH₃CN, and 5 ppbv for CO. There are no systematic biases or obvious anomalies in the simulated frequency distributions. Figure 7 shows scatterplots of simulated versus observed concentrations of HCN and CH₃CN for the ensemble of TRACE-P data. The bulk of the points are strongly correlated, but the correlations are much smaller at high mixing ratios (primarily in the boundary layer), likely because of model errors in the location of plumes. For further evaluation of the simulation we examine vertical distributions and correlations between species.

4.1. Vertical Distributions

The observed and simulated vertical profiles of HCN, CH₃CN, and CO for the ensemble of TRACE-P measurements are shown in Figure 1. The model reproduces the increase of CO mixing ratios and decrease of HCN and CH₃CN mixing ratios from the free troposphere to the MBL. For the background conditions over the NW Pacific (Figure 2), the model reproduces the gradients between the free troposphere and the MBL driven by ocean uptake. Also shown in Figure 1 are the relative contributions from biomass burning and residential coal burning simulated in the model using tagged tracers. Biomass burning accounts for 70-85% of total HCN and 90-95% of total CH₃CN in the free troposphere (above 2 km), while in the boundary layer the fractions are 45% and 75%, respectively. Thus both gases can be used as tracers of biomass burning in the free troposphere, but not in the boundary layer. Simulated and observed mean vertical profiles of HCN and CH₃CN show enhancements at 2-4 and 8-11 km (Figure 1). These correspond to preferential altitudes for frontal and convective outflow of

biomass burning pollution, respectively [Bey *et al.*, 2001b; Liu *et al.*, 2002].

We also compare model results in Figure 8 to aircraft measurements of CH_3CN made over the northern Indian Ocean during the INDOEX campaign [de Laat *et al.*, 2001]. In contrast to TRACE-P, the INDOEX observations show maximum concentrations in the MBL due to boundary layer outflow from the Indian subcontinent associated with the winter-spring monsoon. The model reproduces this enhancement, demonstrating that it is consistent with our representation of the ocean sink.

4.2. Correlations Between HCN, CH_3CN , and CO

Figures 9a and 9b compares the observed and simulated correlations between HCN, CH_3CN , and CO for the ensemble of the TRACE-P data in the free troposphere and the boundary layer. The model reproduces the slopes of the correlations to within their errors both in the biomass burning plumes in the free troposphere (Figure 9a) and in the boundary layer (Figure 9b). Differences in slopes between the free troposphere and the boundary layer reflect the dominant influence of biomass burning in the free troposphere, and the added influence from the urban source (residential coal burning) and the deposition sink in the MBL. The CH_3CN -CO correlation is poor in the boundary layer ($r = 0.29$ in the observations and $r = 0.24$ in the model), while the HCN-CO correlation is strong ($r = 0.85$ in the observations and $r = 0.77$ in the model), reflecting the consistency between the urban and biomass burning emission ratios for HCN/CO but the much lower urban ratios for CH_3CN /CO. The CH_3CN -CO relationship in the boundary layer shows two branches, both in the model and in the observations, reflecting (1) the subsidence of biomass burning influence from the free troposphere (high CH_3CN , low CO), and (2) the urban outflow influence (low CH_3CN , high CO).

4.3. HCN Columns

Spectroscopic measurements of HCN atmospheric columns were made at three Japanese sites during TRACE-P as part of a long-term observational program [Zhao *et al.*, 2000, 2002]. The three sites are Moshiri (44°N , 142°E , 0.28 km altitude), Rikubetsu (43.5°N , 144°E , 0.37 km altitude), and Tsukuba (36°N , 140°E , 0.03 km altitude). The data for the TRACE-P period are shown in Figure 10a (solid circles). The columns at Tsukuba show much higher variation than at Moshiri and Rikubetsu, reflecting the latitudinal difference and the proximity of Tsukuba to the main channel of Asian outflow at 25 - 35°N [Bey *et al.*, 2001b; Liu *et al.*,

this issue]. The increases in the columns from February to April are consistent with the previous observation of a spring-summer HCN column maximum at these sites [Zhao *et al.*, 2000, 2002], and can be explained by the seasonal variation of biomass burning emissions [Li *et al.*, 2000].

Figure 10 compares model results for the TRACE-P period with HCN columns at three Japanese sites. HCN columns were retrieved from solar infrared spectra, and the averaging kernel and *a priori* vertical profile from the retrievals [Zhao *et al.*, 2000, 2002] were applied to the model results. The HCN columns average 4×10^{15} molecules cm^{-2} with high values of $5 - 7 \times 10^{15}$ molecules cm^{-2} , both in the observations and the model, consistent with previous column measurements at northern middle latitudes [Mahieu *et al.*, 1995, 1997; Zhao *et al.*, 2000, 2002]. A scatterplot of simulated versus observed HCN columns is shown in Figure 11. There is strong correlation ($r = 0.82$), reflecting in part the seasonal increase from February to April (Figure 12), and no significant bias.

Multi-year measurements of HCN columns are available from several sites in the northern hemisphere and provide another important test for the model. Figure 11 compares the seasonal variations in these multi-year records, including data from the Japanese sites [Zhao *et al.*, 2000, 2002], Kitt Peak in Arizona [Mahieu *et al.*, 1995], Jungfraujoch in the Swiss Alps [Mahieu *et al.*, 1995, 1997], and Ny Ålesund in Spitsbergen [Li *et al.*, 2000], to model results for 2000. Observations for 1998 were excluded because of anomalous fire influence throughout the northern hemisphere [Rinsland *et al.*, 1999, 2000; Zhao *et al.*, 2002].

The observed spring-summer maxima observed are due to biomass burning in the northern tropics in winter-spring, with a phase lag from meridional transport and augmented by forest fires at middle and high latitudes in summer [Li *et al.*, 2000]. The good simulation of the relative seasonal amplitude constrains the lifetime of HCN and is consistent with the dominance of the ocean sink. The May-July maxima at the Japanese sites are due to fires in Mongolia and northern China that are present in all years of the record Zhao *et al.* [2000, 2002].

5. Global Budgets and Distributions of HCN and CH₃CN

The global atmospheric budgets of HCN and CH₃CN from the model are shown in Table 2. Biomass burning contributes 0.63 TgNyr^{-1} to global HCN and 0.47 TgNyr^{-1} to global CH₃CN, while the corresponding sources from residential coal burning are 0.2 and

0.03 TgNyr⁻¹, respectively. Ocean uptake dominates the loss of both HCN (0.73 TgNyr⁻¹) and CH₃CN (0.36 TgNyr⁻¹), while the sink from reaction with OH is relatively small (0.1 TgNyr⁻¹ for HCN and 0.13 TgNyr⁻¹ for CH₃CN). The resulting tropospheric lifetimes are 5.3 months for HCN and 5.8 months for CH₃CN. Our HCN lifetime is longer, and the implied global source smaller, than the constraints derived in our previous work from the observed relative seasonal variation of HCN columns [Li *et al.*, 2002, Table 1]. However, the present simulation does not seem incompatible with the observed columns (Figure 12).

Figures 13 shows simulated monthly mean concentrations of HCN and CH₃CN in April 2001. Surface concentrations are maximum over biomass burning regions in SE Asia, central America, and central Africa (> 1000 pptv). Surface concentrations are very low (< 50 pptv) at high southern latitudes, reflecting the remoteness from sources and the sink from ocean uptake under our assumption of uniform saturation ratio. Observations in this part of the world are evidently needed. Concentrations in the middle troposphere at 500 hPa are highest over SE Asia and downwind, with a band of relatively high concentrations (> 200 pptv for HCN) at 5-40°N, due to biomass burning influence.

6. Conclusions

We used a global 3-D model analysis of in situ HCN-CH₃CN-CO aircraft observations made over the North Pacific during the TRACE-P aircraft mission (February-April 2001) to improve understanding of the atmospheric budgets of both gases. Vertical gradients of HCN and CH₃CN observed in remote marine air confirm the previous hypotheses of a dominant ocean sink for HCN [Li *et al.*, 2000] and CH₃CN [Hamm and Warneck, 1990]. Following Singh *et al.* [this issue], we used these gradients to derive deposition velocities of 0.13 cm s⁻¹ for both gases, corresponding to saturation ratios of 0.79 for HCN and 0.88 for CH₃CN. These deposition velocities imply lifetimes of three months for HCN(aq)/CN⁻ and 14 months for CH₃CN(aq)/CN⁻ against consumption in the oceanic mixed layer.

Observations from TRACE-P in biomass burning plumes from SE Asia confirmed the importance of biomass burning as a global source for HCN and CH₃CN, while observations in southern California confirmed that automobile exhaust is not a significant source for either gas [Singh *et al.*, this issue]. However, HCN and CH₃CN observed in fresh Chinese urban plumes indicate emissions of anthropogenic origin with a much higher HCN/CH₃CN ratio than from biomass burning. We tentatively attribute these enhancements to residential coal burning.

We use in our global model biomass burning and residential coal burning emission ratios relative to CO of 0.27% and 1.6% respectively for HCN, and 0.2% and 0.25% respectively for CH₃CN, as providing the best fit to the TRACE-P observations within the constraints offered by the HCN-CH₃CN-CO relationships observed in biomass burning and urban plumes. It also offers a good simulation of observed HCN columns, both during TRACE-P and over the multi-year record, and of CH₃CN vertical distributions over the northern Indian Ocean. The resulting simulation captures the observed frequency distributions of HCN and CH₃CN, their vertical distributions, and their correlations with CO.

Our global budgets of HCN and CH₃CN indicate biomass burning sources of 0.63 and 0.47 TgNyr⁻¹, respectively, and residential coal burning sources of 0.2 and 0.03 TgNyr⁻¹, respectively. Ocean uptake is the dominant sink for both; oxidation by OH is an additional minor sink. The resulting tropospheric lifetimes are 5.3 months for HCN and 5.8 months for CH₃CN. The HCN lifetime is longer, and the global source smaller, than in our previous model analysis [Li *et al.*, 2000], but it still provides an unbiased simulation of the relative seasonal cycle of HCN columns which was the main constraint in that previous analysis. The model predicts maximum surface air concentrations of HCN and CH₃CN over biomass burning regions and very low (< 50 pptv) concentrations at high southern latitudes due to ocean uptake under the assumption of a uniform saturation ratio. Observations are evidently needed to test this assumption. Concentrations in the middle troposphere are highest over tropical biomass burning regions and downwind, with a band of high concentrations at 5-40°N, due to biomass burning influence. The dominance of biomass burning as a source of both gases in the free troposphere (70-85% for HCN and 90-95% for CH₃CN) in TRACE-P implies that they can be used there as biomass burning tracers. However, more work is needed to characterize the nature of the Chinese urban source observed in TRACE-P.

Acknowledgments. Discussions with Jennifer Logan and Rose Yevich were very helpful. This work was funded by the Atmospheric Chemistry Program of the U.S. National Science Foundation and by the NASA Global Tropospheric Chemistry Program.

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Figure Captions

Figure 1. Vertical distributions of HCN, CH₃CN, and CO for the ensemble TRACE-P data over the North Pacific. Individual observations are shown as dots. Solid circles, triangles, and horizontal bars represent means, medians, and standard deviations, respectively.

Figure 2. Vertical distributions of HCN and CH₃CN observed in TRACE-P over the North Pacific under background conditions (CO < 120 ppbv and C₂Cl₄ < 10 pptv). Individual observations are shown as dots. Solid circles, triangles, and horizontal bars represent means, medians, and standard deviations, respectively.

Figure 3. Observed HCN (solid circles) and CH₃CN (open circles) correlations with CO in three Chinese urban plumes sampled during TRACE-P (DC-8 flights 9, 12, and 13) below 2 km altitude. Regression lines (RMA method) are also shown.

Figure 4. Observed CH₃CN:HCN correlations in plumes sampled during TRACE-P. Solid circles: sampled at 8-11 km on DC-8 flights 4, 7, 9, 12, 13, 14, 15, 17, 18, and 19; plumes at that altitude were largely devoid of anthropogenic influence [Liu *et al.*, this issue]. Open circles: Chinese urban plumes below 2 km (DC-8 flights 9, 12, and 13). Regression lines are shown.

Figure 5. Observed HCN and CO correlations with Methanol for the ensemble of TRACE-P data. Regression lines are shown.

Figure 6. Frequency distributions of observed (solid line) and simulated (dashed line) mixing ratios of HCN, CH₃CN, and CO for the ensemble TRACE-P data over the North Pacific.

Figure 7. Scatterplots of simulated versus observed mixing ratios of HCN and CH₃CN for the ensemble of TRACE-P data. The 1:1 line is also shown.

Figure 8. Vertical distribution of CH₃CN concentrations over the northern Indian ocean in March 2001. Median aircraft observations from the INDOEX campaign [de Laat *et al.*, 2001] are shown as symbols with standard deviations. Monthly mean model results for the corresponding month and region (0-10°N, 65-80°E) are shown as the solid line.

Figure 9a. Observed and simulated HCN-CH₃CN-CO correlations in the free troposphere (above 2 km) during TRACE-P. Regression lines and correlation coefficients are shown.

Figure 9b. Same as Figure 9a but the boundary layer (below 2 km).

Figure 10. Time series of observed (solid circles) and simulated (open circles) 24-hour average HCN columns (molecules cm⁻²) at the three Japanese stations of Moshiri (44°N, 142°E, 0.28 km altitude), Rikubetsu (43.5°N, 144°E, 0.37 km altitude), and Tsukuba (36°N, 140°E, 0.03 km altitude) during TRACE-P (February-April 2001).

Figure 11. Scatter plot of observed and simulated daily HCN columns (molecules cm⁻²) at Moshiri, Rikubetsu, and Tsukuba in TRACE-P. Values are averages over the three sites. The regression line is calculated without the outlier.

Figure 12. Monthly mean HCN total columns (molecules cm⁻²) retrieved from spectroscopic measurements at Moshiri and Rikubetsu in Japan [Zhao *et al.*, 2000, 2002], Kitt Peak in Arizona [Mahieu *et al.*, 1995], Jungfraujoch in the Swiss Alps [Mahieu *et al.*, 1995, 1997], and Ny Ålesund in Spitsbergen [Li *et al.*, 2000]. Data for the two Japanese sites are monthly means and vertical bars indicate standard deviations. The solid lines represent simulated daily average HCN total columns for 2000.

Figure 13. Simulated monthly mean concentrations of HCN and CH₃CN in surface air and at 500 hPa for April 2001.

Table 1. Global HCN and CH₃CN source estimates (unit: TgNyr⁻¹)

| | Biomass burning | Biofuels | Car exhaust | Residential coal | Biogenic | Total | References |
|--------------------|--------------------|----------|----------------|---------------------|----------|---------|---|
| HCN | | | | | | | |
| | | | | | | 0.2 | <i>Cicerone and Zellner</i> [1983] ^a |
| | 0.37-1.89 | | | | | | <i>Lobert et al.</i> [1990] ^b |
| | 0.64-3.18 | | 0.04 | | | | <i>Lobert et al.</i> [1991] ^c |
| | 0.1-0.3 | | | | | | <i>Holzinger et al.</i> [1999] ^d |
| | 1.4-2.9 | | | | | 1.4-2.9 | <i>Li et al.</i> [2000] ^e |
| | 0.26 | 0.21 | | | | 0.47 | <i>Andreae and Merlet</i> [2001] ^f |
| | 0.7-0.9 | | <0.02 | | 0.2 | 1.1 | <i>Singh et al.</i> [this issue] ^g |
| | 0.63 | | | 0.2 | | 0.83 | this work ^h |
| CH ₃ CN | | | | | | | |
| | 0.27 | | 0.09 | | | 0.36 | <i>Hamm and Warneck</i> [1990] ⁱ |
| | 0.14-0.75 | | | | | | <i>Lobert et al.</i> [1990] ^b |
| | 0.23-1.13 | | | | | | <i>Lobert et al.</i> [1991] ^c |
| | 0.14-0.34 | | 0.01 | | | | <i>Holzinger et al.</i> [1999, 2001] ^d |
| | 0.27 | 0.17 | | | | 0.44 | <i>Andreae and Merlet</i> [2001] ^f |
| | | | | | | 0.44 | <i>de Laat et al.</i> [2001] ^j |
| | 0.4-0.5 | | <0.02 | | 0.2 | 0.7 | <i>Singh et al.</i> [this issue] ^g |
| | 0.47 | | | 0.03 | | 0.5 | this work ^h |

^aSource needed to balance the global sink from reaction of HCN with OH. However, we now know that this reaction makes only a small contribution to the total HCN sink [*Li et al.*, 2000].

^bBased on emission factors per unit nitrogen of $2.42 \pm 1.79\%$ for HCN and $0.95 \pm 0.74\%$ for CH₃CN measured in laboratory biomass burning experiments, and a global biomass burning nitrogen volatilization estimate of 24-57 TgNyr⁻¹ [*Crutzen and Andreae*, 1990]. No detail of the car exhaust source estimate is given.

^cBased on biomass burning molar emission ratios relative to CO of 1.1% for HCN and 0.25% for CH₃CN measured in laboratory biomass burning experiments, combined with a global biomass burning CO source estimate of 480-860 TgCOyr⁻¹.

^dBiomass burning sources based on molar emission ratios (relative to CO) of 0.12% for HCN and 0.13%

for CH_3CN measured in a laboratory biomass burning experiment, combined with a global biomass burning CO source estimate of $210\text{--}550 \text{ Tg CO yr}^{-1}$. The automobile exhaust source of CH_3CN is based on a measured CH_3CN molar emission ratio of 0.06 relative to benzene, combined with an emission ratio of 0.6 g benzene per unit fuel (kg C) consumed in traffic [*de Fre et al.*, 1994] and a global fuel consumption estimate of $2.6 \times 10^3 \text{ Tg C yr}^{-1}$ [*Marland et al.*, 1998].

^eSource needed to balance the HCN loss derived from the relative seasonal amplitude of the column observations, when fitted in a global model simulation (GEOS-CHEM version 3.4) with biomass burning as the only source and ocean uptake as the dominant sink.

^fBiomass burning sources based on HCN molar emission ratios relative to CO of 0.05% for savanna and grassland and 0.15% for tropical and extratropical forests, and CH_3CN molar emission ratios relative to CO of 0.11% for savanna, grassland, and tropical and extratropical forests. These ratios are combined with a global biomass burning CO source estimate of $480 \text{ Tg CO yr}^{-1}$. Biofuel source estimates are based on molar emission ratios relative to CO of 0.20% for HCN and 0.16% for CH_3CN and a global biofuel CO source estimate of $210 \text{ Tg CO yr}^{-1}$.

^gBased on biomass burning and biofuel molar emission ratios relative to CO of 0.27% for HCN and 0.15% for CH_3CN , and a global biomass burning CO source estimate of $700 \text{ Tg CO yr}^{-1}$ [*Andreae and Merlet*, 2001]. Additional biogenic sources of 0.2 Tg N yr^{-1} for both HCN and CH_3CN are invoked to balance estimated global loss rates from ocean uptake.

^hGlobal 3-D model study with biomass burning molar emission ratios relative to CO of 0.26% for HCN and 0.15% for CH_3CN and a global biomass burning CO source estimate of $487 \text{ Tg CO yr}^{-1}$ [*Duncan et al.*, 2002]. The residential coal burning source is invoked to explain enhanced HCN and CH_3CN in Chinese urban plumes. It assumes molar emission ratios relative to CO of 1.4% for HCN and 0.5% for CH_3CN and a global residential coal burning CO source of 30 Tg CO yr^{-1} .

ⁱThe biomass burning source estimate is based on an emission ratio of $2.6 \pm 1.7 \times 10^{-4} \text{ g CH}_3\text{CN g C}^{-1}$ for straw and savannah grass [*Lobert*, 1989] and a global biomass burning carbon source estimate of $3.1 \times 10^3 \text{ Tg C yr}^{-1}$ [*Seiler and Crutzen*, 1980]. The automobile exhaust source estimate is based on a CH_3CN emission factor of 1.3% g/g relative to organic compounds measured in the exhaust of a gasoline-powered engine [*Dulson*, 1978], combined with a global car exhaust source of organic compounds of 21 Tg yr^{-1} .

^jBased on biomass burning and biofuel CH_3CN molar emission ratios (relative to CO) of 0.13% [*Holzinger et al.*, 1999] and a global biomass burning and biofuel CO source estimate of $670 \text{ Tg CO yr}^{-1}$.

Table 2. Atmospheric budgets of HCN and CH₃CN

| | HCN | CH ₃ CN |
|--|------|--------------------|
| Atmospheric burden (Tg N) | 0.43 | 0.28 |
| Atmospheric lifetime (months) | 6.2 | 6.7 |
| Tropospheric burden(Tg N) ^a | 0.37 | 0.24 |
| Tropospheric lifetime (months) | 5.3 | 5.8 |
| Sources (Tg N yr ⁻¹) | | |
| Biomass burning | 0.63 | 0.47 |
| Residential coal burning | 0.2 | 0.03 |
| Sinks (Tg N yr ⁻¹) | | |
| Ocean uptake | 0.73 | 0.36 |
| Reaction with OH | 0.1 | 0.14 |

^aFor the 1000-120 hPa column.

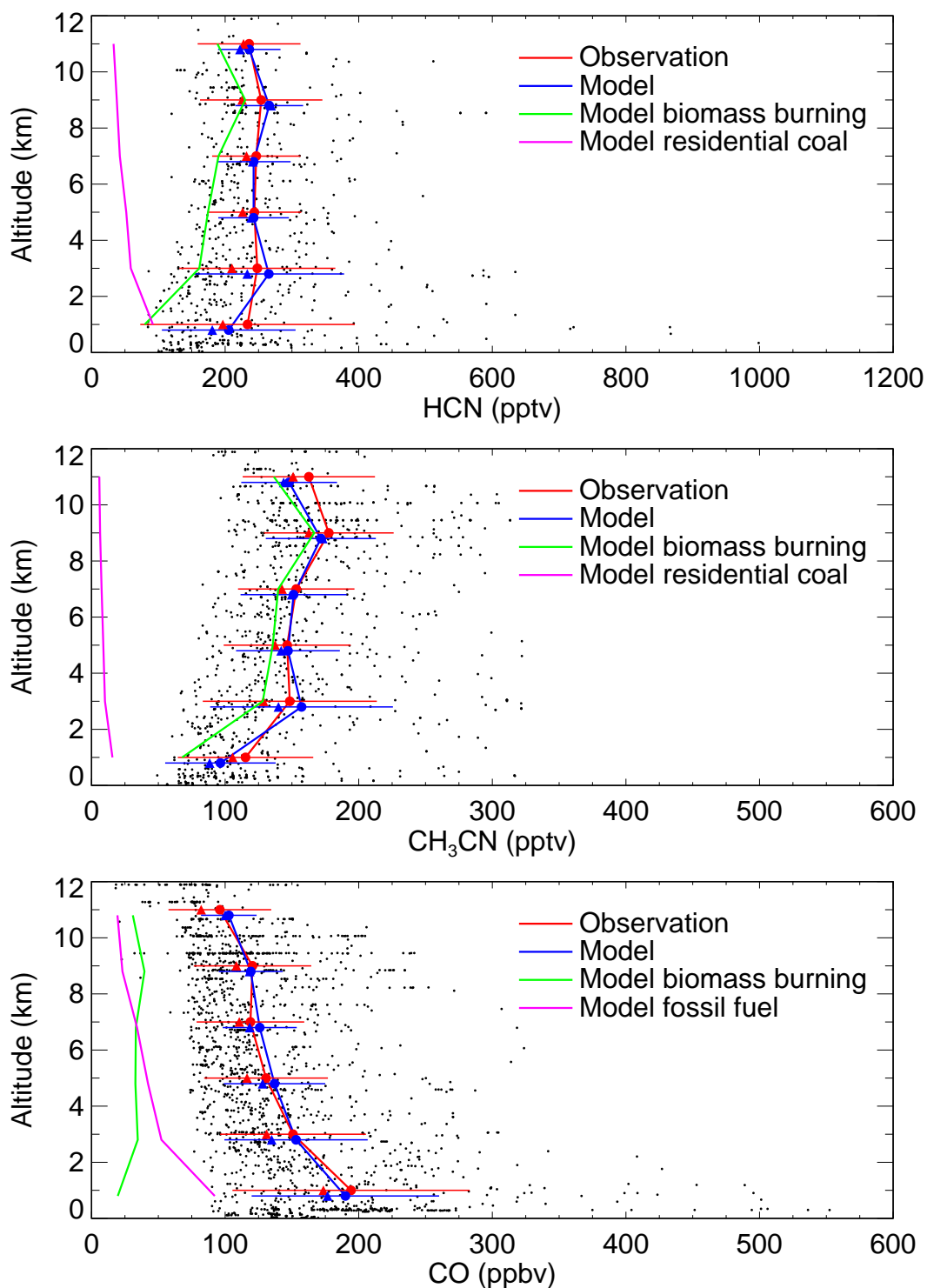


Figure 1. Vertical distributions of HCN, CH₃CN, and CO for the ensemble TRACE-P data over the North Pacific. Individual observations are shown as dots. Solid circles, triangles, and horizontal bars represent means, medians, and standard deviations, respectively.

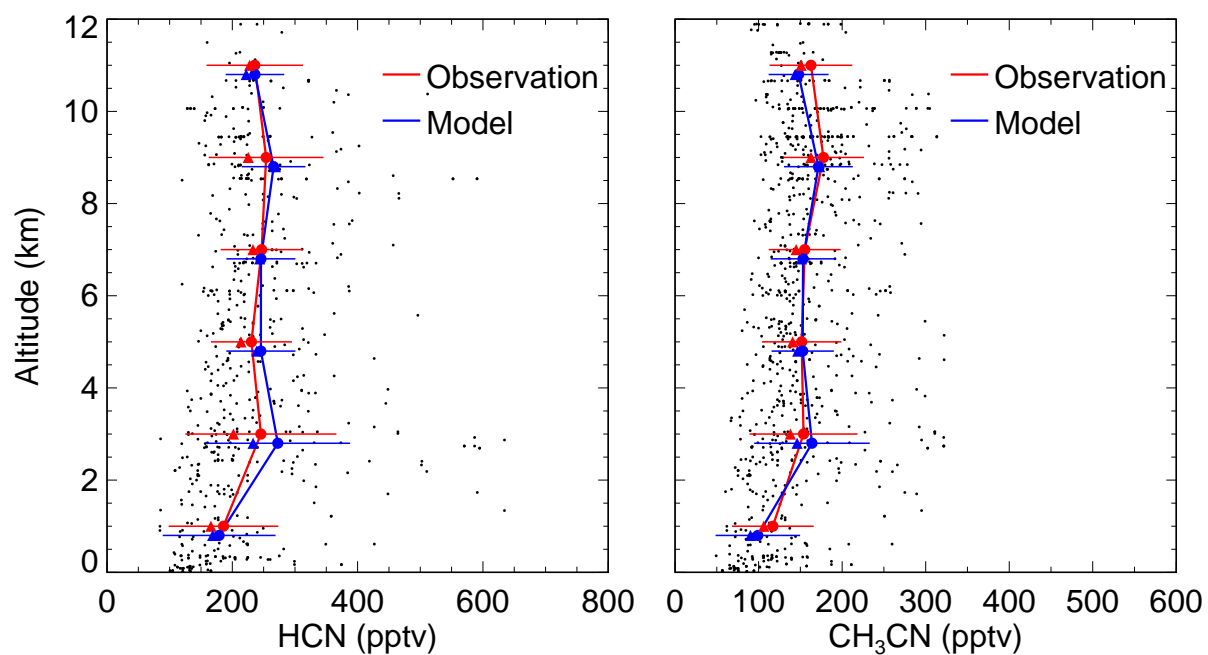


Figure 2. Vertical distributions of HCN and CH₃CN observed in TRACE-P over the North Pacific under background conditions (CO < 120 ppbv and C₂Cl₄ < 10 pptv). Individual observations are shown as dots. Solid circles, triangles, and horizontal bars represent means, medians, and standard deviations, respectively.

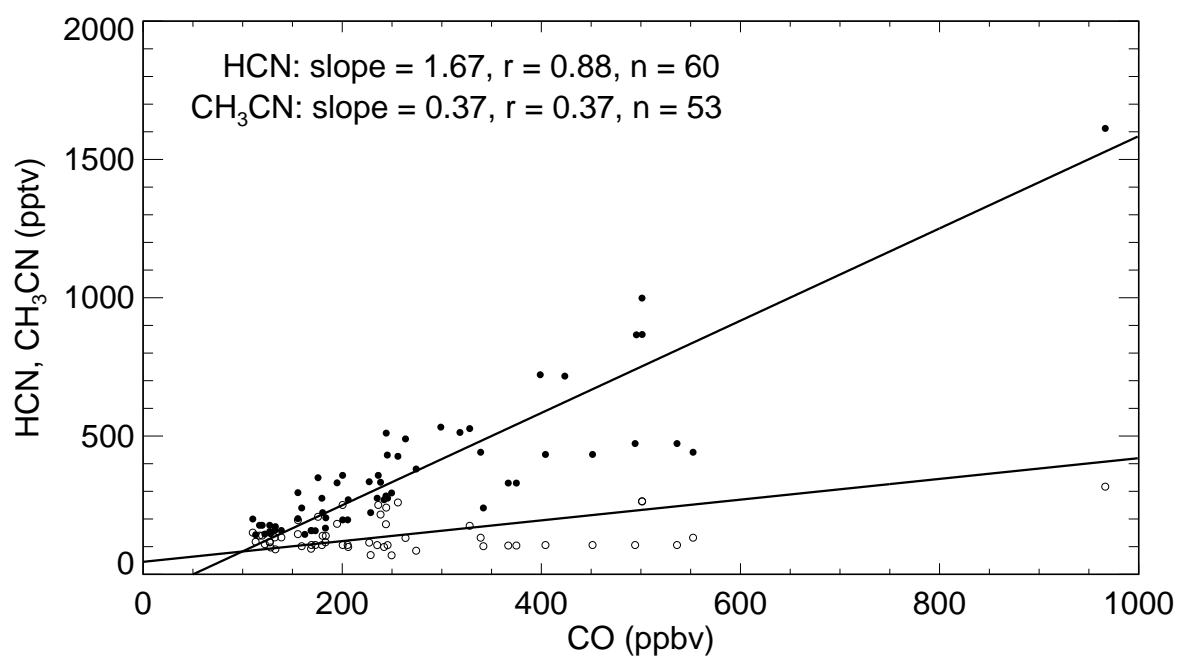


Figure 3. Observed HCN (solid circles) and CH₃CN (open circles) correlations with CO in three Chinese urban plumes sampled during TRACE-P (DC-8 flights 9, 12, and 13) below 2 km altitude. Regression lines (RMA method) are also shown.

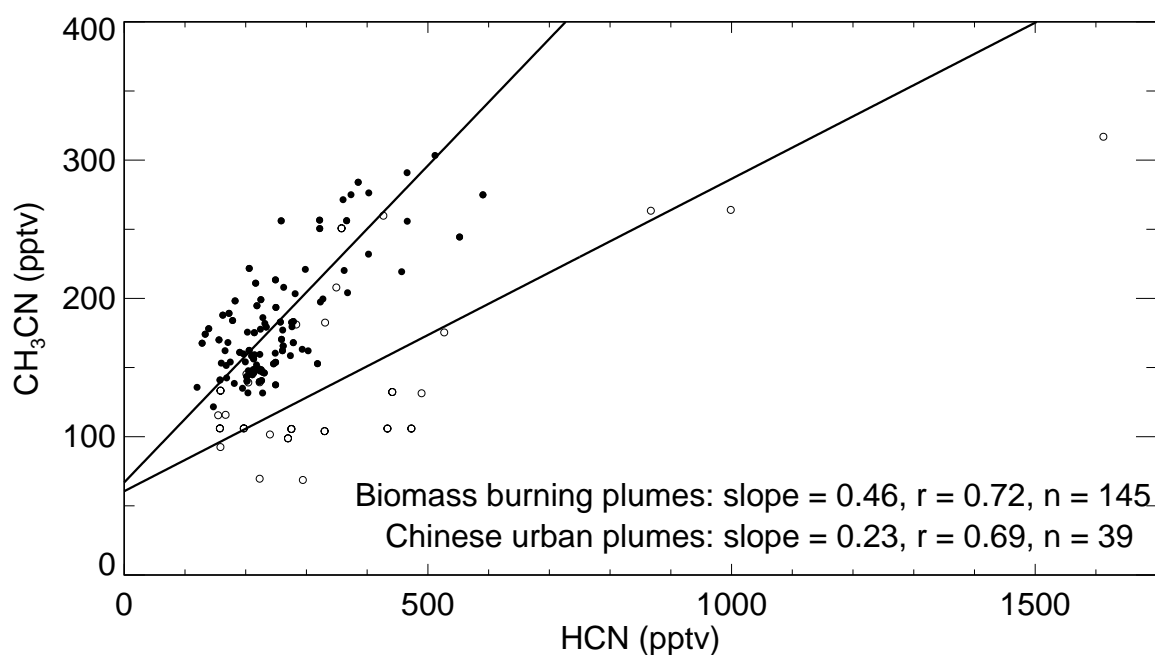


Figure 4. Observed CH₃CN:HCN correlations in plumes sampled during TRACE-P. Solid circles: sampled at 8-11 km on DC-8 flights 4, 7, 9, 12, 13, 14, 15, 17, 18, and 19; plumes at that altitude were largely devoid of anthropogenic influence [Liu *et al.*, this issue]. Open circles: Chinese urban plumes below 2 km (DC-8 flights 9, 12, and 13). Regression lines are shown.

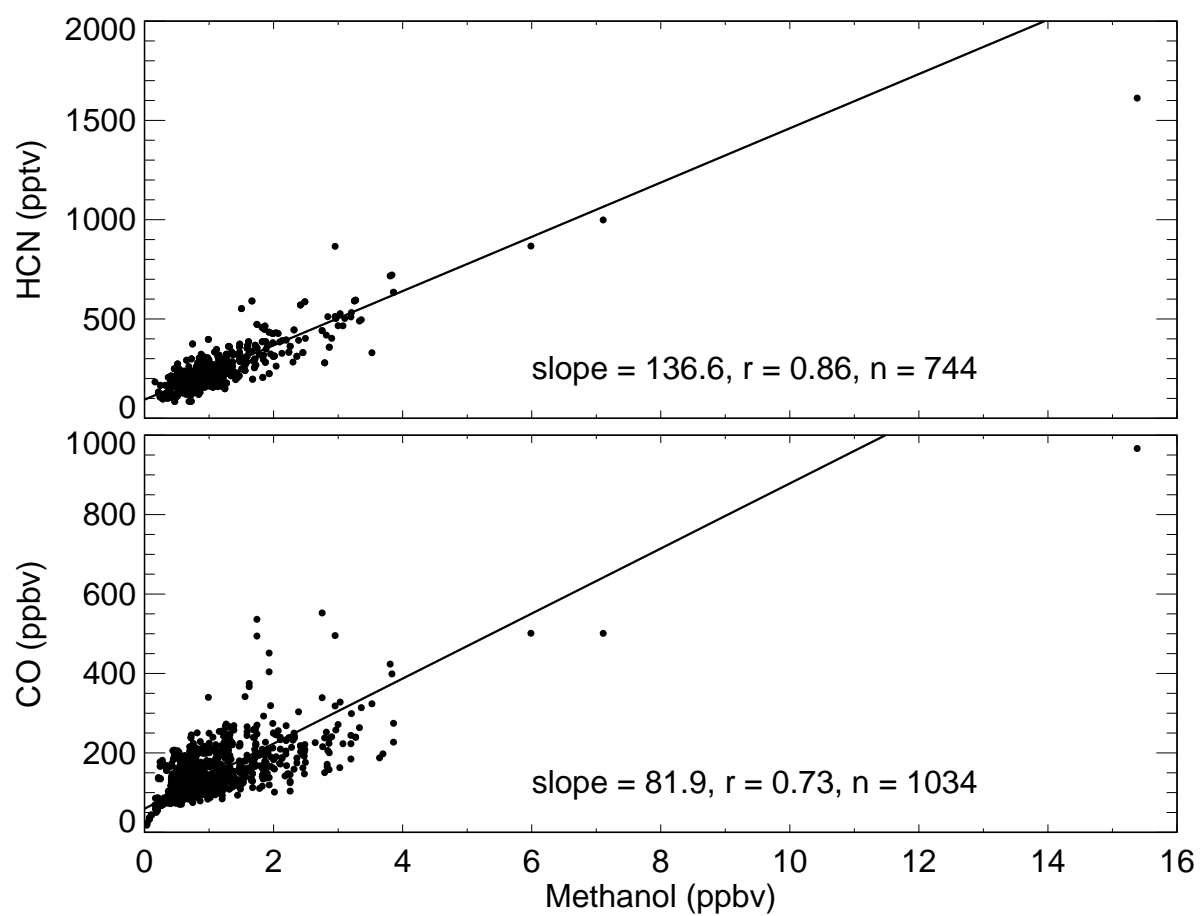


Figure 5. Observed HCN and CO correlations with Methanol for the ensemble of TRACE-P data. Regression lines are shown.

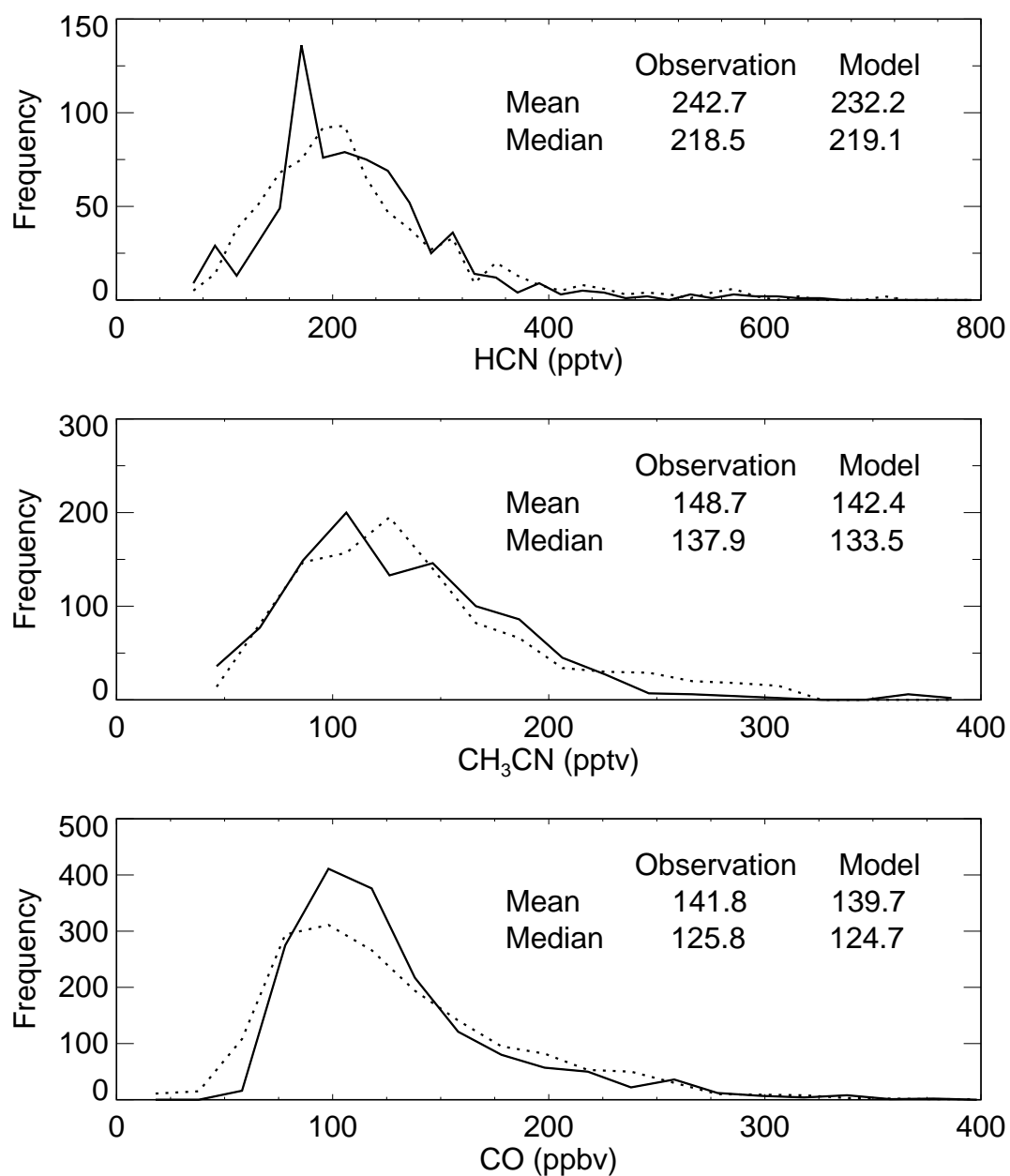


Figure 6. Frequency distributions of observed (solid line) and simulated (dashed line) mixing ratios of HCN, CH₃CN, and CO for the ensemble TRACE-P data over the North Pacific.

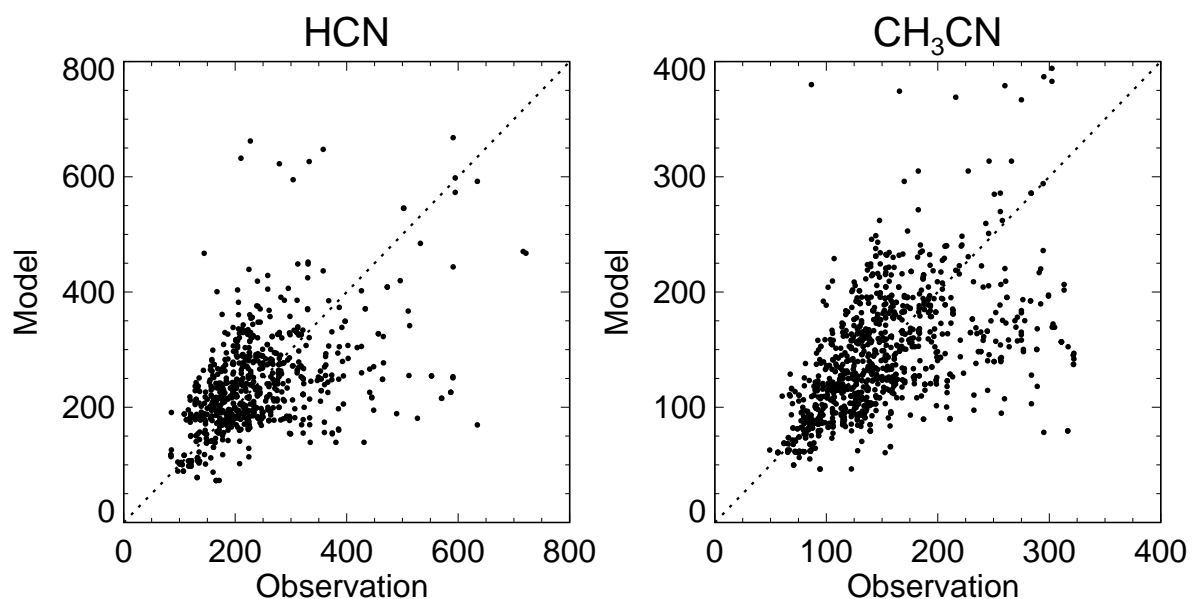


Figure 7. Scatterplots of simulated versus observed mixing ratios of HCN and CH₃CN for the ensemble of TRACE-P data. The 1:1 line is also shown.

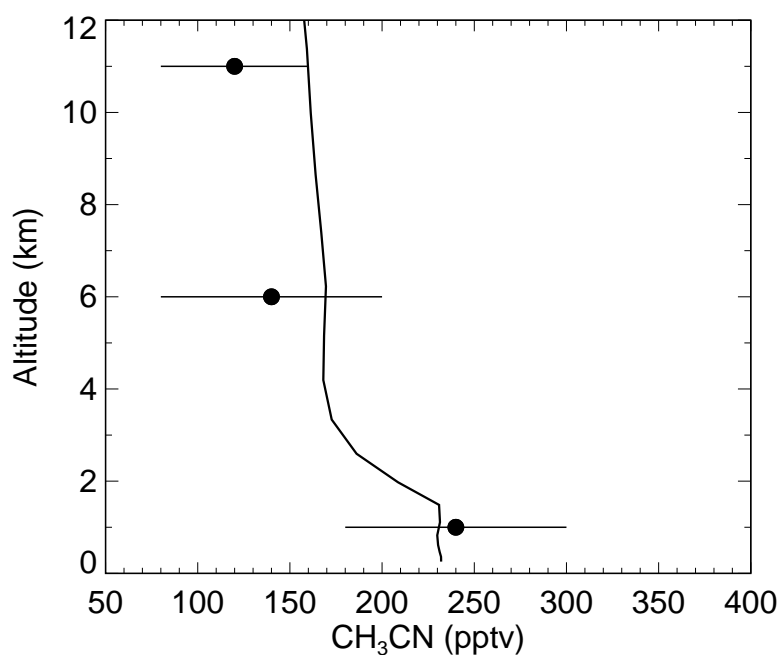


Figure 8. Vertical distribution of CH₃CN concentrations over the northern Indian ocean in March 2001. Median aircraft observations from the INDOEX campaign [*de Laat et al.*, 2001] are shown as symbols with standard deviations. Monthly mean model results for the corresponding month and region (0-10°N, 65-80°E) are shown as the solid line.

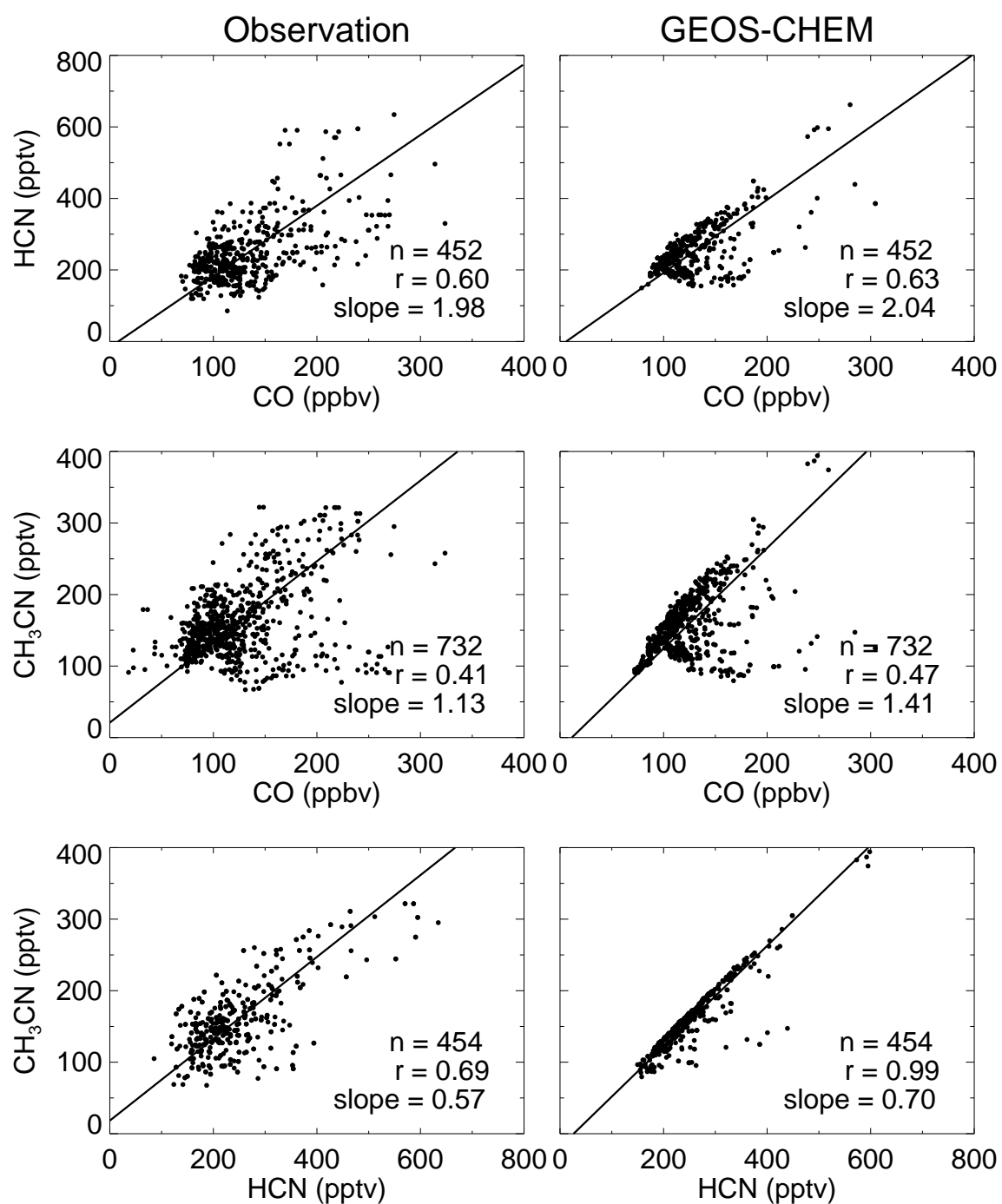


Figure 9a. Observed and simulated HCN-CH₃CN-CO correlations in the free troposphere (above 2 km) during TRACE-P. Regression lines and correlation coefficients are shown.

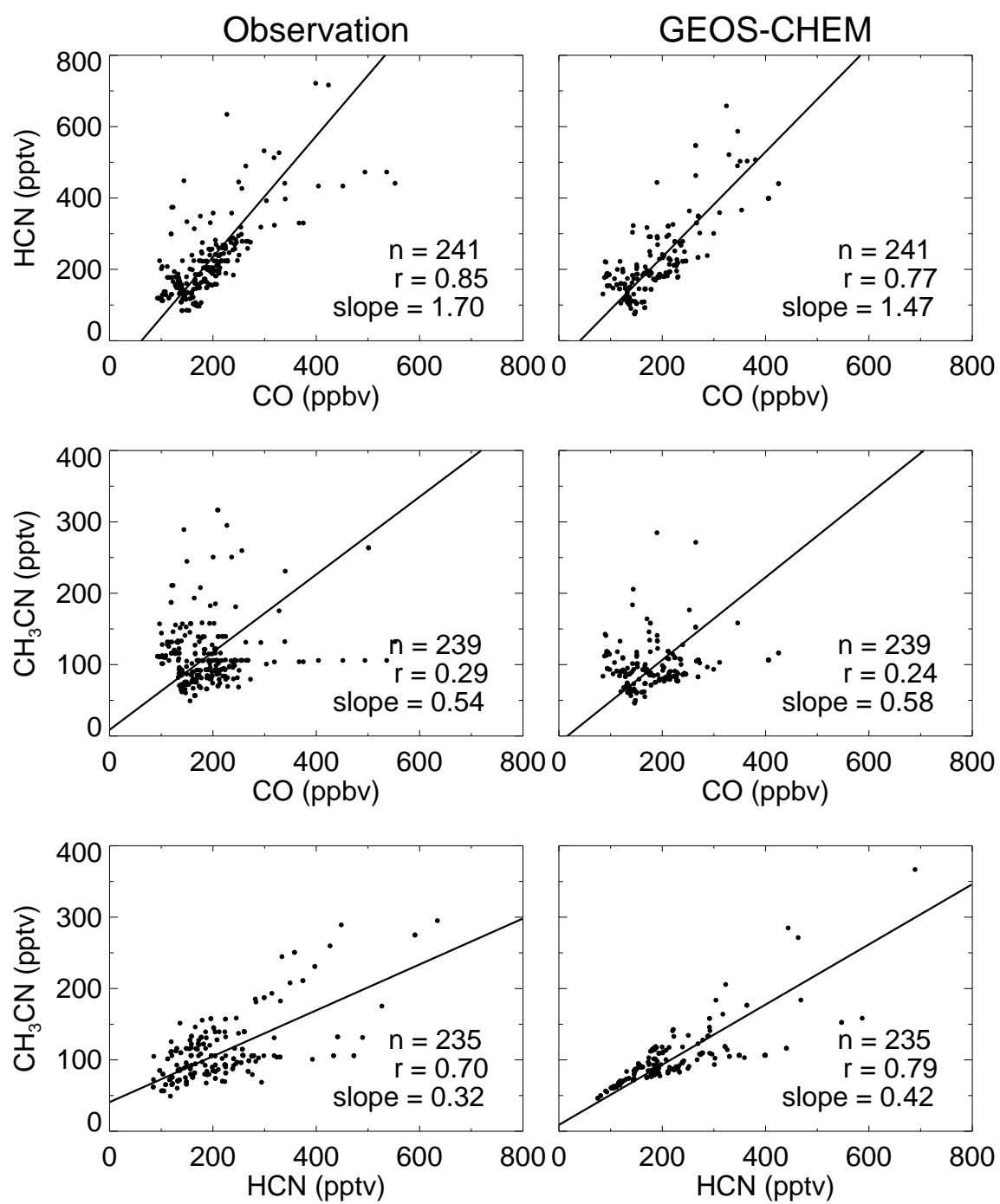


Figure 9b. Same as Figure 9a but the boundary layer (below 2 km).

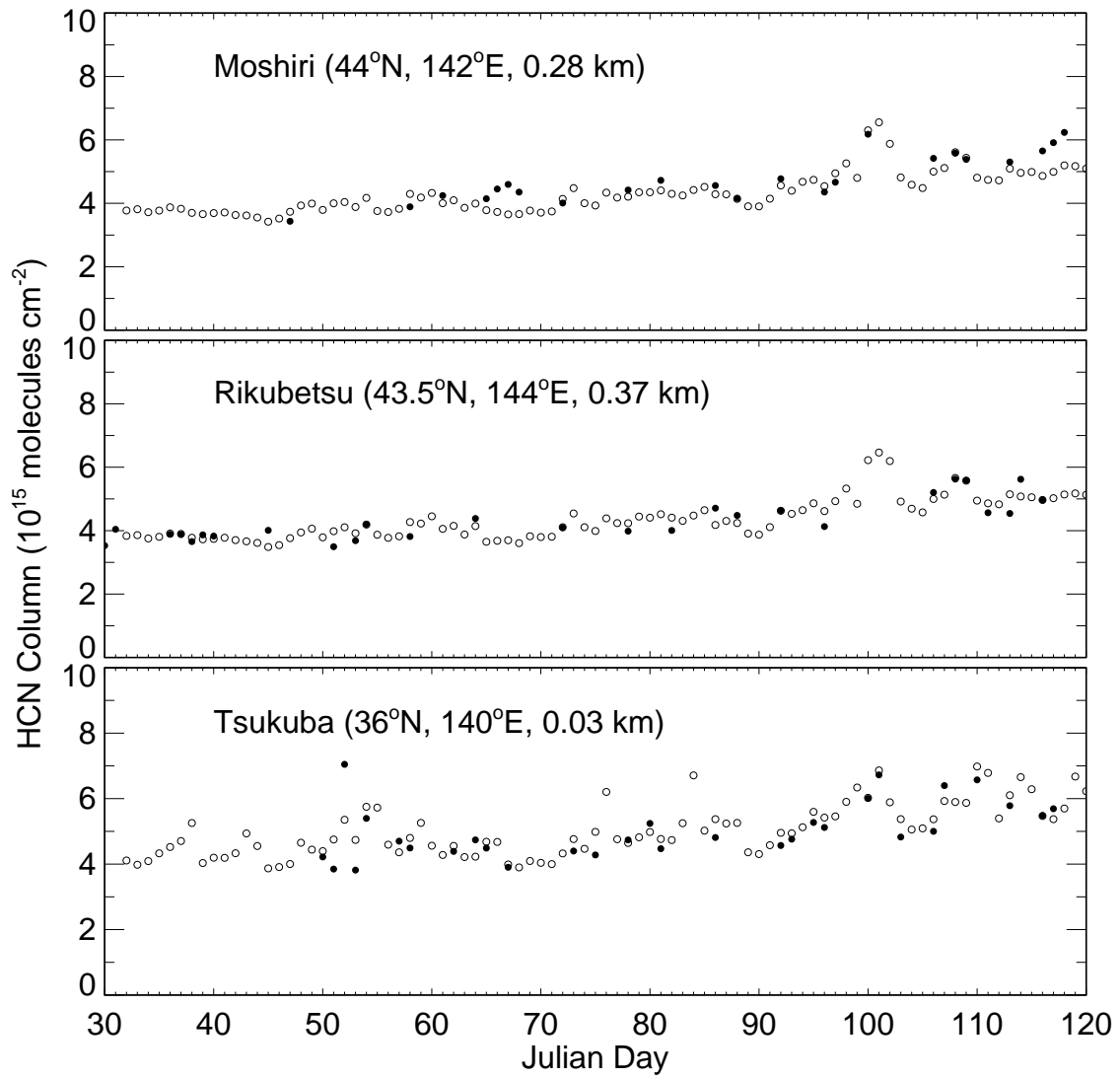


Figure 10. Time series of observed (solid circles) and simulated (open circles) 24-hour average HCN columns (molecules cm^{-2}) at the three Japanese stations of Moshiri (44°N, 142°E, 0.28 km altitude), Rikubetsu (43.5°N, 144°E, 0.37 km altitude), and Tsukuba (36°N, 140°E, 0.03 km altitude) during TRACE-P (February–April 2001).

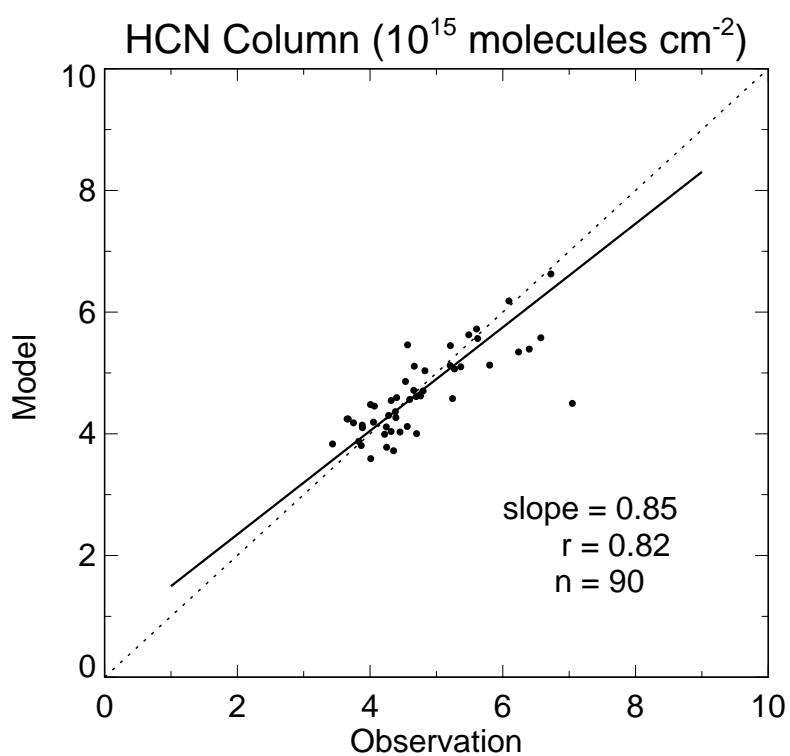


Figure 11. Scatter plot of observed and simulated daily HCN columns (molecules cm^{-2}) at Moshiri, Rikubetsu, and Tsukuba in TRACE-P. Values are averages over the three sites. The regression line is calculated without the outlier.

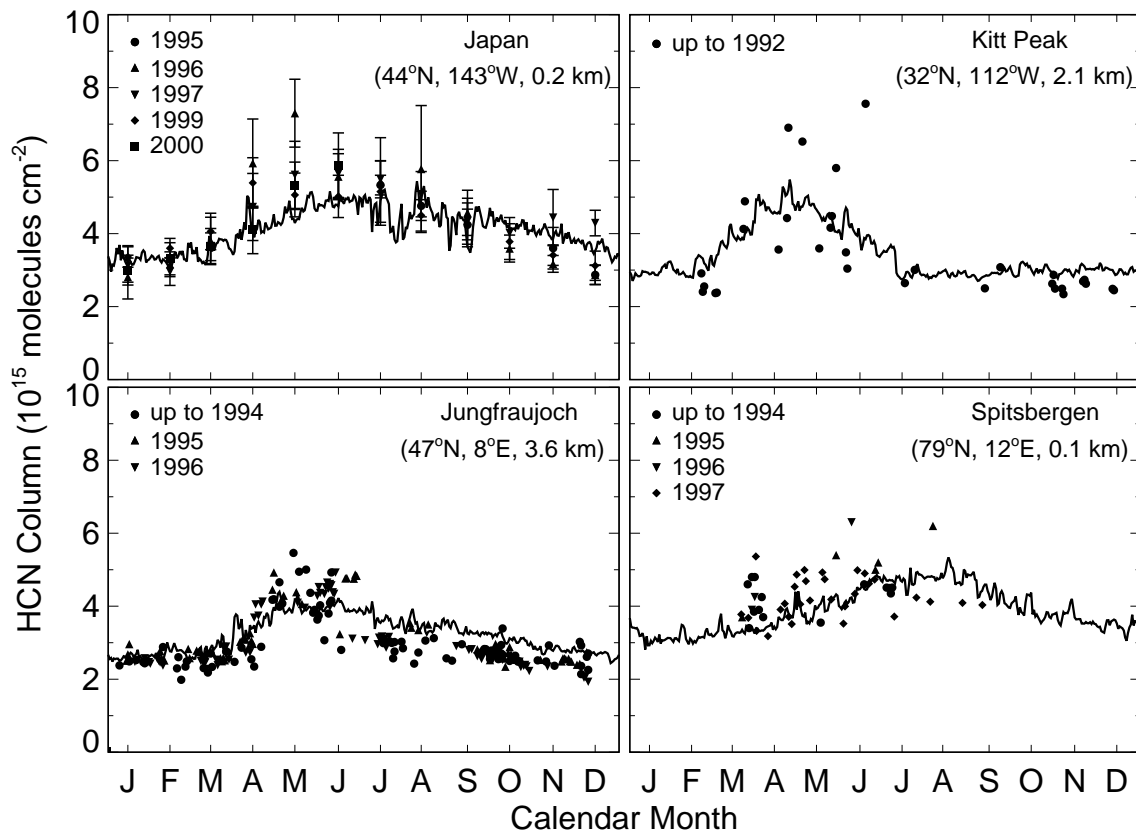


Figure 12. Monthly mean HCN total columns (molecules cm^{-2}) retrieved from spectroscopic measurements at Moshiri and Rikubetsu in Japan [Zhao *et al.*, 2000, 2002], Kitt Peak in Arizona [Mahieu *et al.*, 1995], Jungfraujoch in the Swiss Alps [Mahieu *et al.*, 1995, 1997], and Ny Ålesund in Spitsbergen [Li *et al.*, 2000]. Data for the two Japanese sites are monthly means and vertical bars indicate standard deviations. The solid lines represent simulated daily average HCN total columns for 2000.

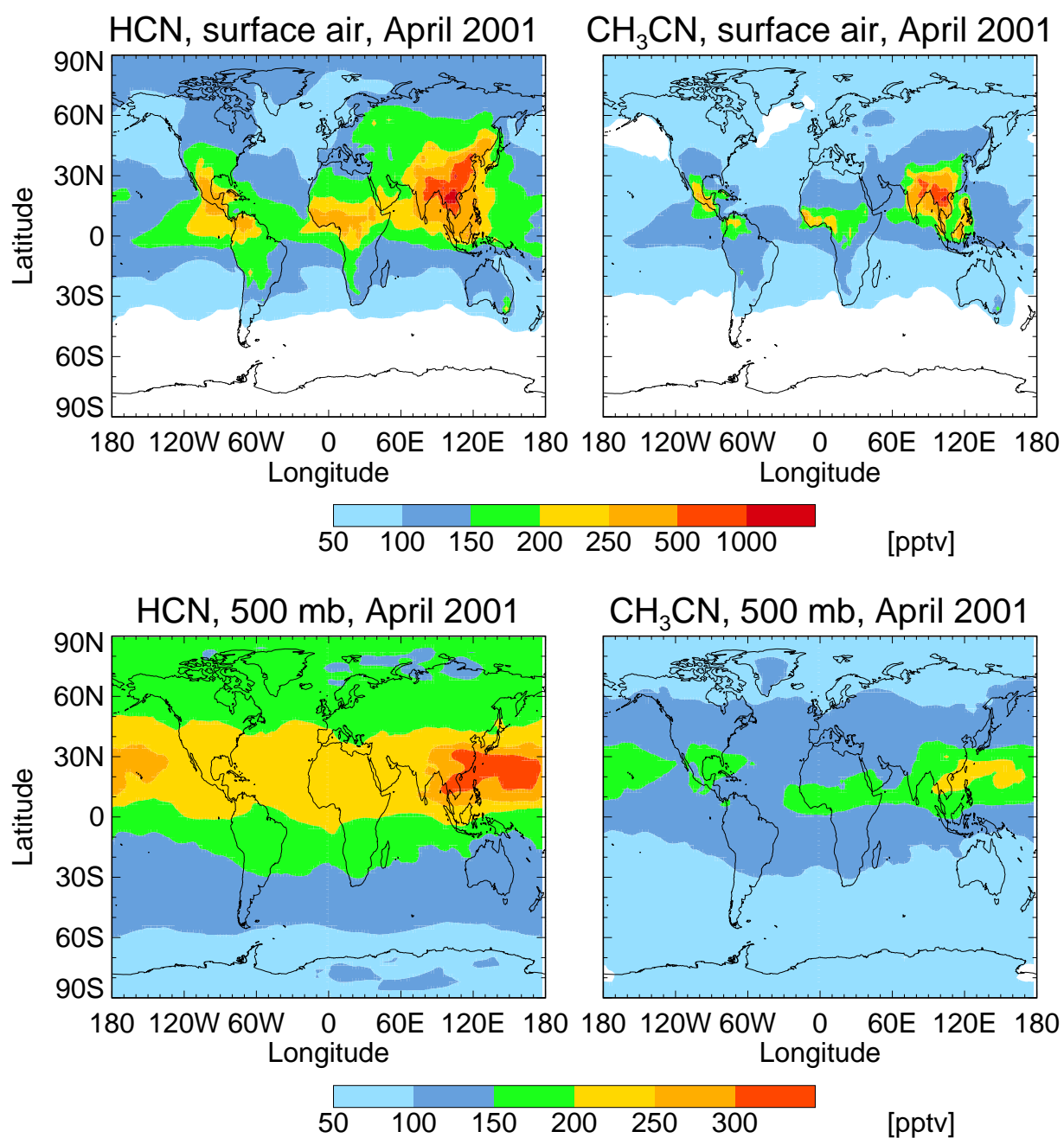


Figure 13. Simulated monthly mean concentrations of HCN and CH₃CN in surface air and at 500 hPa for April 2001.